Sulfate Reducing Bacterial Corrosion of Mild Steel in Liquid and Solid Media

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ABSTRACT

This study was carried out to check the corrosive impact of sulfate-reducing bacteria (SRB) on mild steel coupons in solid and liquid media under two variable nutritional conditions. The bacterial species for the experimental trials was isolated from buried corroded metallic installment and found motile, Gramnegative, non-spore former and identified by 16S rRNA gene sequencing as *Desulfovibrio desulfuricans*. The corrosive impact of SRB on steel coupons was performed in water as liquid medium and the processed clay as the solid medium without and with the provision of Postgate B medium. After 60 days of anaerobic incubation, corrosion rate (CR) and average percent weight loss (APWL) of the steel pieces were calculated. Higher CR and APWL were observed in the liquid medium (water) than in the solid medium (clay). The CR and APWL in water were $249 \pm 2 \ \mu g \ dm^{-2} \ d^{-1}$ and $4.16 \pm 0.25 \ \%$, respectively. In addition, the CR and APWL were higher in the presence of Postgate B medium compared to the medium without it. X-ray diffraction (XRD) and Fourier transform infrared (FTIR) spectroscopic analyses reveals the presence of FeS in the corrosion product. Our findings of this study will be helpful for devising corrosion-protecting strategies.

INTRODUCTION

Metals and their alloys are very important for the construction of buildings and manufacturing of different useful equipment but varying environmental conditions are posing big challenges to metals' stability and becoming the major cause of metallic corrosion (Videla and Herrera, 2005; Ferreira *et al.*, 2007). Microbiologically influenced corrosion (MIC), in this context, is a very complex phenomenon involving interaction between metals, microorganisms, and their environmental conditions. In different environments, MIC deteriorates metals with the involvement of oxidation and reduction

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reactions (Beech and Sunner, 2004; Chugh *et al.*, 2022). In oxidation and reduction reactions, one metal donates electrons, and the other metal accepts electrons and drives the corrosion process. The flow of electrons between these two metal sites determines the CR and severity of metals' damage (Kaesche, 2003; Enning and Garrelfs, 2014).

The estimated damage cost due to MIC is in billion dollars per year (Thompon et al., 2007). In many industries including the gas and oil industry, MIC is a major challenge (Li et al., 2016; Okoro et al., 2016; Liu et al., 2017). A wide variety of microorganisms are involved in the corrosion of metals including bacteria and fungi (Liu et al., 2015). Fungi are the abundantly available eukaryotic microorganisms causing metallic corrosion (Qu et al., 2017; Kadaifçiler et al., 2022; Zhang et al., 2022b). Fungi have the potential to produce organic acids and other corrosive metabolites. These metabolites and organic acids play a crucial role in the corrosion of metals and their alloys (Little and Lee, 2007). Among prokaryotic microbes, acid-producing, iron-oxidizing, sulfate and nitrate-reducing bacterial groups are considered the major contributors causing MIC by colonization on buried gas and oil pipelines (Dang and Lovell, 2015; Qian and Cheng, 2017; Lv et al., 2019). Acid-producing bacteria

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(APB) produce predominately organic acids that cause corrosion of metallic installations (Juzeliunas et al., 2007; Qu et al., 2017; Madirisha et al., 2022). APB can survive under anaerobic environments and initiate the corrosion process (Gu et al., 2018). Nitrate-reducing bacteria such as Bacillus licheniformis and Pseudomonas aeruinosa are also considered the causative agents of MIC (Xu et al., 2013). Iron-oxidizing and sulfate-reducing bacteria (SRB) are the key bacterial groups causing corrosion of materials under aerobic and anaerobic conditions, respectively (Lv et al., 2019). However, among all the corrosion-causing bacterial groups, SRB are considered the key culprits due to their widespread distribution in anoxic environments of aquatic ecosystems and the frequent availability of sulfate there (Hussain et al., 2016; Lv and Du, 2018). Anoxic environments may contain sulfate concentrations up to 28 mM (Pester et al., 2012). SRB consume sulfate as their terminal electron acceptors and generate corrosive sulfide as well as organic acids that cause localized pitting corrosion of metals (Hussain et al., 2016; Zhang et al., 2022a). Various SRB genera (Desulfovibrio, Desulfobulbus, Desulfococcus, Desulfonema and Desulfotomaculum) have been reported to play a vital role in corrosion (Lv and Du, 2018). Keeping in view the wide-ranged environmental distribution and corrosion-causing potential of SRB, the present study was designed for studying the corrosive impact of pure cultured Desulfovibrio desulfuricans on mild steel in water and clay.

MATERIALS AND METHODS

Isolation and molecular identification of SRB

The SRB species was isolated from a wastewater stream (Hudiara Drain) flowing through Mohlanwal, Lahore, Pakistan metallic installations. For the purpose, water (about 100 mL) was collected in a sterile glass container and transported to the laboratory for the isolation of SRB. The bacteria were isolated in deep agar by the dilution to extinction method as described by Postgate (1984) using the Postgate B medium (concentration in g L⁻¹: CaSO₄, 1.0; FeSO₄.7H₂O, 0.5; KH₂PO₄, 0.5; MgSO₄.7H₂SO₄, 2.0; NH₄Cl, 1.0; Ascorbic acid, 0.1; Sodium lactate, 3.5; Thioglycolic acid, 0.1; yeast extract, 1.0).

The bacterial pure culture was characterized phenotypically by motility detection, Gram's, and endospore staining. The bacterial strain was then identified at the molecular level by 16S rRNA gene sequencing. For the purpose, SRB growth was revived in the Postgate B medium. Total genomic DNA from freshly grown bacterial cells was extracted and amplified following Hussain *et al.* (2014) using universal primers [27f (5'-AGAGTTTGATCMTGGCTCAG-3') and 1492r (5'-GGTTACCTTGTTACGACTT-3')]. The amplified gene was then got sequenced through Macrogen, Korea. Neighbor-joining phylogenetic tree of the bacteria isolated in the present study was constructed along with reference sequences recovered from the NCBI GenBank database.

Collection and processing of clay

Clay was collected from the vicinity of UVAS, Ravi Campus, Pattoki and processed in the laboratory for the experimental trial. The collected clay fraction was dried completely in an electric oven (Bioevopeak, DON-18E) at 105 °C till the achievement of constant weight and then sieved to obtain its fine powder (particle size ~ 1 mm). The sieved clay was stored in clean, dry, and air-tightened glass containers till the start of the experimental trial.

Mild-steel coupons

The coupons $(1 \times 1 \times 0.1 \text{ cm})$ were procured from the Department of Metallurgy and Materials Engineering, University of the Punjab, Lahore, Pakistan, and degreased with analytical grade acetone (Cat No. 10014, Merck, Germany) as previously described by Bano (2008). The coupons were then polished with polishing paper (Cat No. WD240, Caswell, Taiwan). After cleaning, the coupons were washed with ethanol (Cat No. 100983, Merck, Germany) followed by cleansing with distilled water, and dried completely at 80 °C for 10 min in an electric oven. The initial weights (g) of coupons were measured by using an electric balance after cooling (Zuo and Wood, 2004). The coupons were then autoclaved and buried under different experimental conditions.

Experimental trial

The experimental trial for studying the corrosive impact of SRB on steel coupons was conducted in moist clay and water with and without the provision of Postgate B medium. For studying corrosion in clay in the presence of nutrients, 30 g clay was mixed with 10 mL of Postgate B medium. Similarly, for studying corrosion in clay in the absence of nutrients, 30 g clay was mixed with 10 mL of distilled water. In other experimental sets, only water (30 mL) was used instead of clay with and without the provision of Postgate B medium. All the experiments were conducted in sterile serum bottles (SSB) of 30-mL capacity. Initially, in all the experiments, pH was set between 7.0 and 7.5. The processed and autoclaved coupons were then inserted in the solid media vertically and left horizontally in the liquid media. Half vials of both sets (A and B) were inoculated with revived SRB cultures [5 % (v/v) ~ 1.8×10^6 C.F.U. mL⁻¹], while the remaining half vials were kept un-inoculated (control). Oxygen diffusion in the vials was stopped by sealing SSB with butyl rubber stoppers (Cat No. 32-A2, Sunshine, China) and aluminum crimp seals (Cat No. SU27227-U, Supelco, Germany). The sealed SSB were then incubated at 30 °C for 60 days.

Measurement of CR and APWL

At end of the experimental period, coupons were recovered from the clay under aseptic conditions. Corrosion products and other contaminants were removed from the coupons using Clark's solution (Sb₂O₃, 20 g; SnCl₂, 50 g; HCl, 774 mL) following Angeles-Chavez *et al.* (2001). Contaminant-free coupons were rinsed with distilled water, dried completely, and weighed to check the weight loss. The rate of corrosion (μ g dm⁻² d⁻¹) was then estimated by following the given formula (Majumdar *et al.*, 1999):

$$C = \frac{W1 - W2}{AT}$$

Where, C represents CR (μ g dm⁻²d⁻¹), while W1 and W2 are weights of coupons before and after the harvesting, respectively. A represents the surface area of the coupons and the experimental period (days) is represented by T.

XRD and FTIR spectroscopy analyses of the corrosion product

Fine powder of the corrosion product collected after crushing was processed for XRD analysis for checking the existence and nature of oxides and sulfides. A semicomputer-controlled XRD (RIGAKU, Model DMAX/A) analysis was performed to determine the composition of the corroded material.

The corroded product was then observed with FTIR (Nicolet AVATAR360, Nicolet, Madison, AL, USA) for the quantitative analysis of organic groups produced by SRB during growth and consequent corrosion of the coupons.

Statistical analysis

All the experiments were carried out in triplicates. The data were analyzed by one-way ANOVA using Minitab 19. Mean differences were considered significant at P < 0.05.

RESULTS

The present study reported SRB-induced CR and APWL in clay and water in the absence (A) and presence (B) of some mineral nutrients. For this purpose, SRB species were isolated from a wastewater stream and found motile, Gram-negative, non-spore former and identified by 16S rRNA gene sequencing as *Desulfovibrio desulfuricans*. A comparison of sequence homology with known sequences is shown in Figure 1 as basis for identification. On average, CR and APWL appeared maximally in the experimental coupons of both experimental sets A and B. The CR and APWL were significantly higher in water as compared to



Fig. 1. Neighbour joining phylogenetic tree of *Desulfovibrio desulfuricans* isolated in the present study along with reference sequences recovered from the NCBI GenBank database. The current tree was derived from the analysis of 16S rRNA sequences. Bootstrap support values at each node are indicated by numbers (percentage of 1,000 replicates). Values above 70% are displayed only. The bar represents 0.02 substitutions per nucleotide position. The bacterial isolate from the current study is denoted by green. *Staphylococcus aureus* is used as an outgroup.





CR and APWL in the experimental set A CR and APWL in clay

The CR and APWL in clay appeared almost the same in the control vials as compared to the experimental vials. The CR and APWL in the experimental vials were 43 \pm 2 μg dm⁻² d⁻¹ and 0.75\pm0.25 %, respectively, while the corresponding figures for CR and APWL in the control vials appeared as 46±2 and 0.74±0.11, respectively (Table I). The impact of the SRB remained masked in this set of experiments.

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clay (Fig. 2). The CR and APWL in sets A and B appeared as:

| Index | With postgate B medium | | | | Without postgate B medium | | | |
|-------|------------------------|----------------------|--------------------|----------------------|---------------------------|----------------------|--------------------|----------------------|
| | Clay | | Water | | Clay | | Water | |
| | Inoculated n= 3 | Uninoculated n= 3 | Inoculated n= 3 | Uninoculated n= 3 | Inoculated n= 3 | Uninoculated n= 3 | Inoculated n= 3 | Uninoculated n= 3 |
| CR | 108 ± 3 | 62 ± 2 | 249 ± 2 | 166 ± 2 | 43 ± 2 | 46 ± 2 | 68 ± 2 | 55 ± 2 |
| APWL | 1.88 ± 0.29 | 1.68 ± 0.11 | 4.16 ± 0.25 | 2.83 ± 0.14 | 0.75 ± 0.25 | 0.76 ± 0.11 | 1.22 ± 0.45 | 0.98 ± 0.19 |

Table I. Corrosion rate (CR, µg dm⁻² d⁻¹) and average percent weight loss (APWL, %) of mild steel coupons in clay and water under different experimental conditions.

CR and APWL in water

The CR and APWL in water appeared amazingly higher in the experimental vials as compared to the control vials. The CR and APWL in the experimental vials were $68 \pm 2 \ \mu g \ dm^{-2} \ d^{-1}$ and $1.22 \pm 0.45 \ \%$, respectively, while the corresponding figures for CR and APWL in the control vials appeared as 55 ± 2 and 0.98 ± 0.19 , respectively (Table I). The significant impact of SRB on the experimental coupons was observed in this set of experiments.

CR and APWL in the experimental set B CR and APWL in clay

The CR and APWL in clay appeared slightly higher in the experimental vials as compared to the control vials. The CR and APWL in the experimental vials were 108 ± 3 μ g dm⁻² d⁻¹ and 1.88 ± 0.29 %, respectively, while the corresponding figures for CR and APWL in the control vials appeared as 62 ± 2 and 1.68 ± 0.11 , respectively (Table I). The impact of SRB remained masked in this set of experiments.



Fig. 3. XRD analysis of the corrosion product showing higher peak of iron sulfide.

CR and APWL in water

The CR and APWL in water were significantly higher

in the experimental vials as compared to the control vials. The CR and APWL in the experimental vials were $249 \pm 2 \ \mu g \ dm^{-2} \ d^{-1}$ and $4.16 \pm 0.25 \ \%$, respectively, while the corresponding figures for CR and APWL in the control vials appeared as 166 ± 2 and 2.83 ± 0.14 , respectively (Table I). The significant impact of SRB was observed in this set of experiments.

XRD and FTIR spectroscopy analyses of the corrosion product

Analyses of the corrosion product depicted the presence of FeS in all the inoculated groups, while different iron oxidation forms were appeared in the uninoculated groups (Figs. 3 and 4).



Fig. 4. FTIR analysis of the corrosion product depicting significant presence of metal sulfide.

DISCUSSION

In the current study, the corrosive impact of SRB on mild steel coupons was studied in fluid and solid media. For this purpose, a motile, Gram-negative, and non-spore former SRB species was isolated from a wastewater stream and identified by 16S rRNA gene sequencing as *D. desulfuricans*. SRB have been studied extensively and considered one of the key causative agents of corrosion of buried metals and their alloys (Muyzer and Stams, 2008; Yuan et al., 2013; Guan et al., 2016; Xu et al., 2016).

In this study, two major sets of experiments were performed with and without the addition of nutrients of the Postgate B medium. In experimental set B, the addition of nutrients manifested higher CR and APWL values. Numerous SRB genera (*Desulfobulbus*, *Desulfococcus*, *Desulfonema*, *Desulfotomaculum*, and *Desulfovibrio*) have been reported to play a vital role in corrosion due to the abundant availability of sulfate in anoxic environments (Lv and Du, 2018; Tran *et al.*, 2021). In the presence of nutrients, SRB grew well and generate more H_2S and thus caused more corrosion. It is well known that nascent H_2S acidifies water and causes pitting corrosion of metals. Furthermore, it enhances the corrosive action when combines with O₂ and CO₂ (Asmara, 2018).

Higher CR and APWL values were observed in water when compared with that of clay. The CR and APWL in water were $249\pm2 \ \mu g \ dm^{-2} \ d^{-1}$ and $4.16\pm0.25 \ \%$, respectively, while the corresponding figures for CR and APWL in clay were 108 ± 3 and 1.88 ± 0.27 , respectively. The isolated SRB species for the current study appeared motile. The motile bacteria are more likely to show their impact in liquid media. Motile bacteria may penetrate the pores present on the metal surface and thus result in more severe corrosion. Similar findings have been reported by Jimenez-Lopez *et al.* (2008) while studying the advantageous motility of *Myxococcus xanthus* in porous material thus increasing the consolidation efficiency.

XRD and FTIR spectra showed the presence of FeS in all the inoculated groups. It has been reported that SRB use sulfate as their terminal electron acceptors and a variety of organic substrates as their electron donors and generate H_2S (Hussain *et al.*, 2016). The biogenic H_2S acidifies the moisture contents that cause the pitting corrosion of metals (Butt and Mohsin, 2022; Qin *et al.*, 2022; Xu *et al.*, 2022). In our study, the presence of FeS in the inoculated groups confirmed the production of H_2S and consequent corrosion of the steel coupons.

CONCLUSION

The present study reveals the corrosive impact of SRB on mild steel coupons in clay and water in the absence (A) and presence (B) of some mineral nutrients. On average, CR and APWL appeared maximally in the experimental coupons of both experimental sets A and B. The CR and APWL were significantly higher in water as compared to clay while depicting the corrosive potential of SRB. Future studies on the corrosion behaviour of SRB while utilizing various other electron acceptors and donors are needed to explore more facts on SRB-based corrosion compatible with various geochemical conditions.

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IRB approval

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Statement of conflict of interest

The authors have declared no conflict of interest.

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